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Geostatistical investigation into the temporal evolution of spatial structure in a shallow water table

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1 Introduction

Shallow water tables are highly variable in both time and space. This variability creates difficulty in predicting how water tables respond to rainfall events and where and when saturated areas occur when the water table rises to the surface. This is troublesome because the position of the water table can determine which hydrologic pathways are active. Regions with high water tables can promote the occurrence of saturated areas leading to overland flow. On shallow soils characterized by highly conductive topsoil underlain by dense subsoil, and in regions where the ground water is close to the soil surface, runoff can be generated from regions that are or become saturated during rainfall events (e.g., Dunne

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and Black, 1970a, b; Steenhuis et al., 1995). Rainwater easily permeates these soils and, by-and-large, runs laterally as subsurface flow on top of the restrictive layer down-slope to accumulate in converging areas making these regions prone to saturation. More water can be added to these regions via direct rainfall or exfiltration from higher on the hillslope. When these regions saturate they produce runoff that is commonly termed saturation excess overland flow. How (i.e., exfiltration of converging subsurface flow or direct rainfall onto saturated areas) and what (i.e., old or new water) water finds its way to these regions is not fully understood making the variability in physical patterns of saturated areas difficult to monitor and predict (Burns, 2002; McDonnell, 2003; Walter et al., 2005). The difficulty in capturing the dynamics of these saturated regions stems from their non-linearity in both space and time exhibited among rain events and seasons. Due to this variability, researchers have coined the term variable source area (VSA) to describe these areas (e.g., Dunne and Black, 1970a, b; Hewlett and Hibbert, 1967; Dunne et al., 1975). While important in a pure hydrology perspective (i.e., predicting runoff amounts, peak timing in hydrographs), representing the spatial and temporal nature of VSAs is quintessential to modeling and managing contaminant flow pathways in natural environments. As observed by Grayson et al. (2002), there has been an increased focus in current research on spatial variability to account for where contaminants come from and where to invest financial resources to improve water quality. Although the concept has been around for well over a quarter of a century, it is obvious that the formation of VSAs and how they influence water quality is still a topic of interest for hydrologists.

Repeatedly, the call for better distributed data to aid in understanding hydrological processes, especially for data to identify processes controlling the formation of VSAs, has gone out (Hillel, 1986; Klemes, 1986; Hornberger and Boyer, 1995; Agnew et al., 2006). New methods of collecting and interpreting spatially distributed data to characterize VSAs have become available. Snap shots of soil moisture using various remote sensing techniques (Choudhury, 1991; Engman, 1991; Blyth, 1993; Verhoest et al., 1998; Troch et al., 2000) and field measurements (Western and Grayson, 1998; Mohanty et al., 2000; Meyles et al., 2003; Walker et al., 2001; Wilson et al., 2004) have been used to locate regions concentrating water. These sampling techniques, however, may not be applicable for all field sites. For example, the extremely effective and increasingly popular technique incorporating time domain reflectometry (TDR) sensors mounted to an all-terrain vehicle (Tyndale-Biscoe et al., 1998; Western and Grayson, 1998) is limited by field accessibility. This type of sampling may not be an option for field sites with large biota (e.g., trees, shrubs, corn), extreme geology (e.g., steep slopes, boulders, large gullies), or excessive amounts of surface water (e.g., ephemeral streams, saturated source areas). Satellite remote sensing techniques have their own difficulties, not the least of which include signal inter-

pretation, limited coverage, and low temporal and spatial resolution. While these methods are powerful, they are often too temporally sparse (i.e., low frequency of sampling) to capture the spatial evolution of VSAs. High temporal resolution measurements of water table depth are becoming readily available due to inexpensive, self-contained, water level data loggers (e.g., TruTrack, Inc). Loggers of this type can be employed to monitor depth to water table from the field scale up to the watershed scale. The ability to capture short term changes in the water table depths makes it possible to observe the effects of storm events. Since the position of the water table during storm events is crucial to VSAs, these measurement techniques provide information about where the runoff is being produced. However, few techniques are available to easily summarize this enormous mass of data. Here we will show how geostatistics lend themselves naturally to characterize spatial structure with semivariograms and kriged interpolation among points to obtain realistic spatial patterns of water table heights and saturated areas.

Geostatistical analysis most commonly uses semivariograms to define the variance between two observations as a function of the distance separating them. The main semivariogram parameters are the nugget, the sill, and the range. If a stable sill exists, the process is assumed stationary and the sill can be thought of as the variance between two points separated by a large distance. The range is the measure of the spatial continuity and the maximum distance over which spatial correlation affects the variable of interest. The nugget represents the variance between two close measurements. The nugget value is attributed to variance occurring at scales smaller than the sample spacing and to the inherent sampling device error. Within the realm of semivariogram techniques, indicator semivariograms provide a method to capture extreme values (Journel, 1983). Indicator semivariograms have been used to assess risk of contamination in various constituents such as heavy metals (Webster and Oliver, 1989; Smith et al., 1993; Goovaerts and Journel, 1995) and assess uncertainty in soil properties (McKenna, 1998; Pachepsky and Acock, 1998; Goovaerts, 2001). In their most basic form, indicator semivariograms treat data as a binary indicator with respect to a threshold value (i.e., 1 if threshold is exceeded; 0 if threshold is not exceeded). Both indicator semivariograms and standard semivariograms describe spatial structure by representing variability between observation points. This spatial structure can, in turn, be used to interpolate between observation points using kriging. Kriging provides a way to interpolate and visualize spatial patterns based on observations. More complete discussions of semivariograms, and the associated kriging, along with many possible derivatives in algorithms and methodology are provided in Goovaerts (1997), Deutsch and Journel (1992), and Chilès and Delfiner (1999).

Kriging of various forms has been used to interpolate maps of potentiometric surfaces from water table data (Delhomme, 1978; Neuman and Jacobsen, 1984; ASCE, 1990). The goal

of most research of this nature is how to best interpolate discrete spatial observations into full coverage. To this end, some research has looked at using existing information about the landscape to supplement point observations of the water table. Hoeksema et al. (1990) supplemented well data with elevation in mapping of a phreatic surface using a co-kriging approach. More recently, Desbarats et al. (2002) used kriging with external drift incorporating the TOPMODEL topographic index of Beven and Kirkby (1979) to interpolate water table elevations. Their results showed that predictions made accounting for the traditional topographic index resulted in non-physical water table behavior or higher than observed fluctuations in groundwater in regions of sparse observations. Lyon et al. (2006) used indicator kriging (IK) to incorporate soft data developed using logistic regression. “Soft” data are local information that is a proxy to the variable of interest and need not relate directly (Goovaerts, 1997) as opposed to “hard” data which are actual measurements of the variable of interest. Lyon et al. (2006) were able to improve interpolations for low antecedent rainfall condition rain events using pre-event water table positions as a predictor of saturation. The analysis of Lyon et al. (2006) required information about the pre-event depth to water table that may not be available and cannot be extended beyond the boundaries of the study site. Also, the study made observations on only large storm events and did not look at how spatial structure of the water table changed through time.

This research looked at the spatial and temporal evolution of the shallow water table in the near stream region of a head-water catchment. The position of this shallow water table was directly related to the formation of saturated source areas. Our goal was to characterize both short time interval and long time interval variations, to thus better understand event-based and seasonal hydrologic responses, in the spatial structure of the shallow water table using semivariogram analysis. This type of geostatistical analysis is capable of representing large amounts of data easily. Depth to ground water was measured at 44 locations for 5 min intervals from March 2004 through August 2004. These data were used to develop indicator semivariograms on a small-time interval, event basis and probability of exceeding the time-variable median depth to water table on a long-time interval, seasonal basis. Both the event and seasonal influences can be incorporated into a kriging interpolation to visualize the spatial patterns occurring in the shallow water table on the hillslope. This incorporation provides a manner to supplement spatial observations based on limited, discrete observations using an understanding of the hydrological processes operating on the hillslope. Also, this analysis provides a utility to represent the variability of the shallow water table that affects the formation of saturated regions in both time and space. This representation gives insight to the dominant hydrological patterns in terms of runoff generation at the hillslope scale. The correct characterization and representation of such patterns is essential for hydrologists interested in predicting water

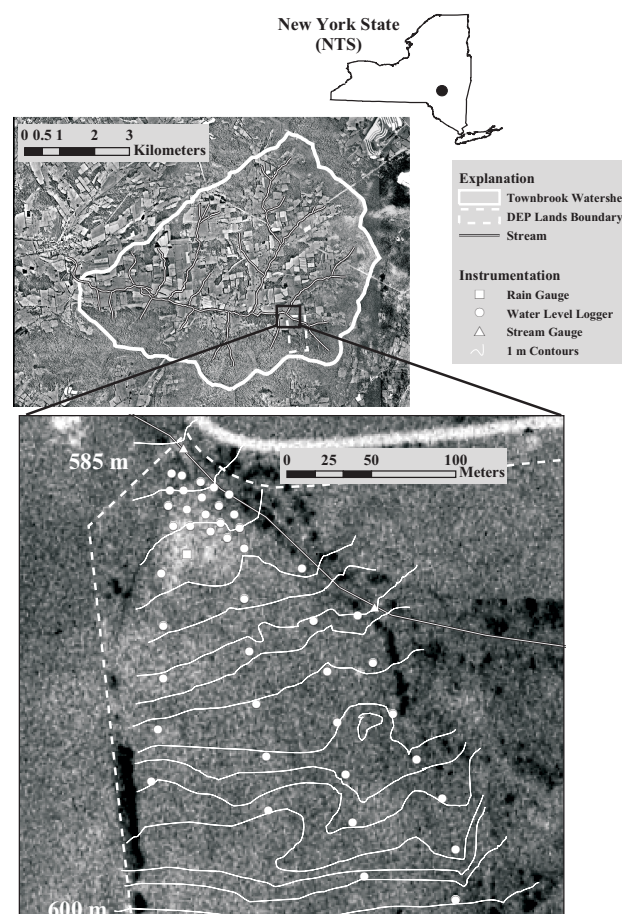


Fig. 1. Location of study site at Townbrook research watershed in Catskill Mountain region of New York State showing positions of sampling locations containing water level loggers (circles), stream gauges (triangles), and rain gauge (square).

movement and chemical transport from the landscape to the stream.

2 Site description and data

The 2.44 ha study site on New York State Department of Environmental Protection (DEP) owned lands is part of a 2 km² sub-watershed located in the southwest corner of the 37 km² Townbrook watershed in the Catskill Mountain region of New York State (Fig. 1). The landuse on the study site is uniformly grass/shrub with forested regions upslope (south) the study area. A survey of more than 200 points was conducted to supplement the existing 10 m digital elevation model (DEM) and derive 1-m interval contours for identifying small-scale topographic features. The study site covers the near stream region ~120 m along the stream (bordering the northern side of the study site) and ~180 m upslope (south) from the stream and elevation varied from 585 m to 600 m above mean sea level with slopes varied from 0° to 8°.

Soil Survey Geographic Database (SSURGO) soil maps were used to determine soil types and properties. Two soil types dominate the study site. The northern (down slope) half of the study site consisted of approximately 30 cm deep gravely silt loam. The southern (up slope) half of the study site consisted of approximately 56 cm deep silt loam. The soil is underlain by a restrictive fractured bedrock layer. These shallow soils were typified by a higher hydraulic conductivity (1.4×10^{-5} m/s) in the surface material and a lower hydraulic conductivity (1.4×10^{-6} m/s) in deeper layers.

At 44 measurement locations piezometers were instrumented for continuous monitoring depth to water table. The water levels in the upper 30 cm of the soil were recorded using WT-HR 500 capacitance probes manufactured by Tru-Track, Inc, New Zealand. Levels were recorded at 5-min intervals and averaged over 15-min intervals for the study period from 10 March 2004 to 22 August 2004. The location of the piezometers followed approximately two grid systems. The first consisted of 20 loggers on a 10×10 m grid near the stream (northern end) of the study site. In addition, 24 loggers were located on a large spacing 30×40 m grid to record water table levels upslope from the stream. A few capacitance probes failed for some periods to record data and need to be repaired, recalibrated, or replaced. During these periods, the sampling location was removed from the data set and assigned a “no data” value and not used in the analysis. At most, two sampling locations from the 44 sampling locations were assigned “no data” values at any given time. A tipping bucket rain gauge with data logger was set on the site to record rainfall amounts. Also, two water level loggers were placed in the stream above and below the study site to gauge the runoff from during the sampling period. These water level loggers recorded the stream stage and were converted to flow using rating curves developed for the stream at both locations. Each rating curve was based on seven current-meter discharge measurements; 16% of all stream height observations required extrapolation beyond the highest known point of the rating curve. Runoff from the hillslope was calculated as the difference in flow downstream and upstream of the study site with negative values during low flow conditions removed. This was reasonable since there was only little catchment area contributing from the other side of the stream for this stream segment (Fig. 1). Rain data and stream data were not available for the last two weeks of the study period (from 6 August).

3 Methods

3.1 Indicator variables

Water table observations were transformed to indicator variables to give information about water table positions deeper than detectable by the piezometers. Indicator variables, which constitute a non-linear transformation, allow for a

more comprehensive structural analysis and are more robust to outlier values (Journel, 1983). In this way, indicator approaches allow for greater spatial correlation of extreme values (Journel and Alabert, 1989; Rubin and Journel, 1991). Although they may not be appropriate for identification of connectivity in some spatial patterns, semivariograms based on indicator variables may provide additional information over traditional, measurement-based semivariograms for data clustering in space (Western et al., 1998a).

3.1.1 Short time intervals

To give a proportional number of observations above and below the threshold, the time-variant median depth to water table at each 15-min interval was used as the threshold. The time-variant median provides the best defined, with greatest range of continuity and confidence for spatially sparse data indicator semivariograms (Journel, 1983). Indicator variables were, thus, defined as:

$$I_i(z_c(t)) = \begin{cases} 1 & \text{if } z_i(t) \leq z_c(t) \\ 0 & \text{if } z_i(t) > z_c(t) \end{cases} \quad (1)$$

where, $I_i(z_c)$ is the indicator value at sampling location i , $z_i(t)$ is the measured depth to water table at sampling location i [cm] at a certain point in time t , and $z_c(t)$ is the median depth to water table [cm] at the same time t . The time-variable threshold ensured that there were equal numbers of zeros and ones in the data set at any time step. With a constant threshold, the number of ones would be time-variable, which would cause artifacts in the geostatistical analysis. It should be noted that a one did not indicate a wet location but rather a location that was wetter than 50% of the wells. The sets of indicator variables for each 15-min time step were used to characterize spatial structure, on a short time interval, to describe event-based changes in the shallow water table.

3.1.2 Long time intervals

Long time interval spatial structure at the study site was evaluated by combining the 15-min data into monthly intervals. Monthly intervals were selected because of their ability to capture the seasonal variability of hydrologically active areas for this region (Walter et al., 2001; Agnew et al., 2006). For each month (March through August), the frequency of the water table at a sampling location exceeding the time-variable median water table (i.e., how often was the water table at a certain location among the 50% wettest locations) was computed to give a probability of exceeding the threshold. This frequency also describes the prior probability of exceeding the threshold used later for the development of soft data.

3.2 Semivariogram generation

Semivariograms were constructed for both the short time interval and the long time interval observations using the

semivariance, $\gamma_s(h)$, at a lag, h :

$$\gamma_s(h) = \frac{1}{2N(h)} \sum_{(i,j)} (Y_i(z) - Y_j(z))^2 \quad (2)$$

where, N is the number of pairs, $Y_i(z)$ and $Y_j(z)$ are the variable of interest at i and j , respectively, with summation over pairs (i, j) . For the short time interval, the variable of interest was the indicator values at points i and j instead of measured values. For long time interval, the variable of interest was the probability of exceeding the threshold at i and j . Plotting the average semivariance for pairs grouped by separation distance or grouped into “bins” defined by a “lag” distance in semivariogram nomenclature against the average bin distance, sample semivariograms were created with Eq. (2) to relate distance between sampling location and semivariance. The sample semivariograms were calculated using 10 bins with lags of 15 m. The number of bins and lag distance were selected using the rule of thumb that the number of bins multiplied by the lag distance should be approximately half the maximum separation distance (288 m for this site). For the short time interval, the semivariance was standardized with the variance of the observations to lower scatter around the sill. The sample semivariograms were then represented using a fitted equation or “model”. The exponential and spherical models, both of which are widely used, were investigated because they offer adequate representation of lags within the range of the sample semivariograms. The exponential model (Eq. 3) was selected for the remainder of this analysis because it better fit the sample semivariograms using a method of weighted least squares (Cressie, 1985):

$$\gamma_e(h) = \sigma_0^2 + \left(\sigma_\infty^2 - \sigma_0^2 \right) \left(1 - e^{-\frac{h}{\lambda}} \right) \quad (3)$$

$\gamma_e(h)$ is the fitted semivariogram, σ_0^2 is the nugget, σ_∞^2 is the sill, and λ is the correlation length. This model reaches its sill asymptotically with the range (i.e., maximum distance over which spatial correlation affects the variable of interest) defined as 3λ .

Thus, for the short time interval data, indicator semivariograms based on indicator values defined with Eq. (1) were created and for the long time interval data, traditional semivariograms were created from the probability of exceeding the time-variable threshold. Using an automated fitting procedure programmed in Matlab v7r14 (The Mathworks, Inc., 2004) exponential models for both the short time interval indicator semivariograms and long time interval semivariograms were created. Since anisotropy was found to be minimal for the study site (Lyon et al., 2006), only omnidirectional semivariograms were used in this study. The parameters of these models describe the spatial structure of the shallow water table and were compared to measured runoff and surface saturation on the hillslope. For this study, saturation was considered when the depth to water table at a sampling location was less than 5 cm, i.e., close to or at the soil surface. The area representing each sampling location that sat-

urates was determined using Thiessen polygons to compute the portion of hillslope saturating.

3.3 Indicator kriging interpolations

3.3.1 Generating hard and soft data

Equation (1) was used to create hard data (i.e., indicator variables) from the short time interval data. To interpolation between sampling locations (for this and all subsequent interpolations), ordinary kriging was performed using the Geostatistical Analyst extension available in ESRI[®] ArcMap[™] v9 (ESRI, Inc., 2004). When using indicator variables, the resulting IK is the probability of exceeding the defined threshold. This indicator kriging using hard data creates a map to visualize spatial patterns of saturation on the hillslope predicted using the short time interval observations.

A major advantage of the IK approach is its ability to account for soft data (Deutsch and Journel, 1992). With this in mind, the second interpolation method for this study was IK coupling hard data with soft data. Soft data can relate prior probabilities about the indicator variables to auxiliary information, such as existing geographic conditions (e.g. soil map, topography). Agnew et al. (2006) demonstrated that the soil topographic wetness index (*STWI*) was a good predictor of saturation for this watershed based on a 30-year modeling simulation. To develop soft data for this study site, the relationship between the prior probability (i.e., monthly frequency a sampling location exceeding the time-variable median water table) and the *STWI* was established. *STWI* is defined as:

$$STWI = \ln \left(\frac{a}{\tan \beta D \hat{K}_s} \right) \quad (4)$$

where a is the area of the upslope watershed per unit contour length [m], $\tan \beta$ is the local slope, D is the soil depth [m] and \hat{K}_s is the mean saturated hydraulic conductivity [m/day]. Values for a and $\tan \beta$ were determined for the study site using the D_∞ algorithm of Tarboton (1997); D and \hat{K}_s were taken from SSURGO soil distribution maps for the study site.

The *STWI* values from each sampling location were categorized into unit intervals (i.e., sampling locations with *STWI* values between 8 and 9 in the first category, between 9 and 10 in the next, etc.) and the average *STWI* was evaluated for each interval. This resulted in six total intervals. The average prior probability for exceeding the median water table was also computed for each interval. A linear function relating the two was created such that:

$$P = mx_{STWI} + b \quad (5)$$

where P the prior probability of exceeding the threshold, x_{STWI} is the average *STWI* for each unit interval, and m and b are the slope and intercept, respectively. This defined prior probability at locations with no observations and was used to create a spatially continuous prior probability map based on

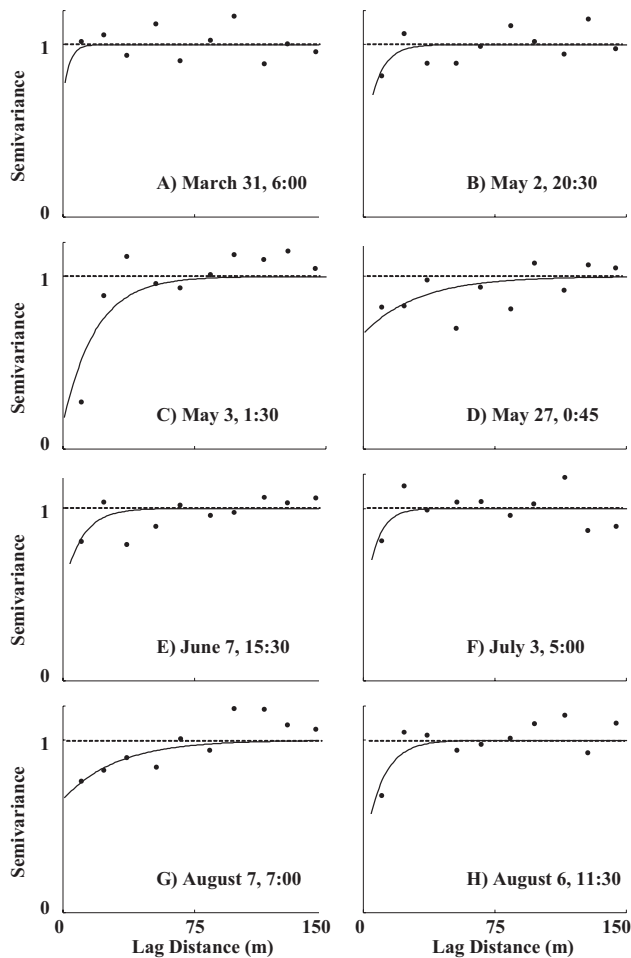


Fig. 2. Typical indicator semivariograms from the short time interval data for the study site for (A) 31 March, 06:00, (B) 2 May, 20:30, (C) 3 May, 01:30, (D) 27 May, 00:45, (E) 7 June, 15:30, (F) 30 June, 04:30, (G) 26 July, 05:30, and (H) 6 August, 11:30. The symbols are the normalized sample indicator semivariogram and the curves are the fitted exponential models.

STWI. This map is soft data based on seasonal trends for the hillslope.

3.3.2 Combining hard and soft data

Residuals were evaluated between the hard data available at sampling locations and the soft data, i.e., *STWI*, map. These residuals were then interpolated and merged with the soft data to incorporate prior probability. This is consistent with the method for incorporating soft data given in Goovaerts (1997). For comparison, the interpolation made using hard data alone (or traditional IK) and the interpolation combining hard and soft data were conducted on data from the six rainfall events causing the highest median water tables for the spring period (March through May).

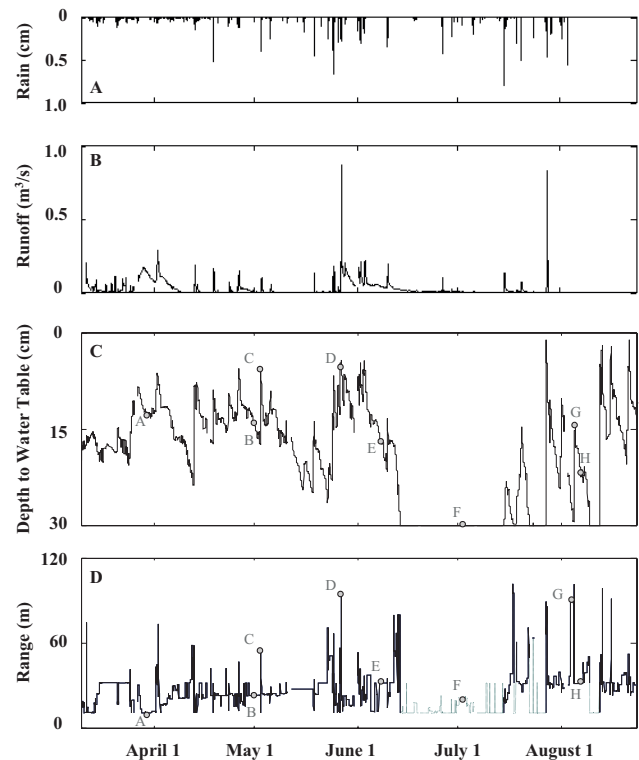


Fig. 3. Short time interval measurements for the study site from 10 March through 22 August of (A) rainfall [cm], (B) runoff [m^3/s], and (C) median depth to water table [cm]. For each indicator semivariogram, the (D) range parameter from exponential model [m]. Circles on (C) and (D) show when in time the indicator semivariograms in Fig. 2 occur and the portion of (D) indicated with a dashed line is where number of sampling sites below minimum detection level is greater than half all sampling sites.

4 Results

4.1 Short time interval

Exponential models were fitted to the indicator semivariograms for various median water tables and at various times in the sampling period (Fig. 2). Many of the indicator semivariograms had a well-defined sill and identifiable ranges. These indicator semivariograms provide information about the spatial structure of the shallow water table for snapshots in time; however, they provide no information about the evolution of the shallow water table with time. To look at this evolution along with changes in rainfall and runoff at the hillslope, time series were created over the sampling period. Low intensity storms were more frequent during the first half of the study period (March through mid-May) while high intensity storms occur in the second half (Fig. 3A). Peaks in runoff coincided with the rainfall events with the large rainfall events producing most runoff from the study site (Fig. 3B). The two largest runoff events occurred on 26 May and 27 July after periods of high antecedent rainfall (1-day

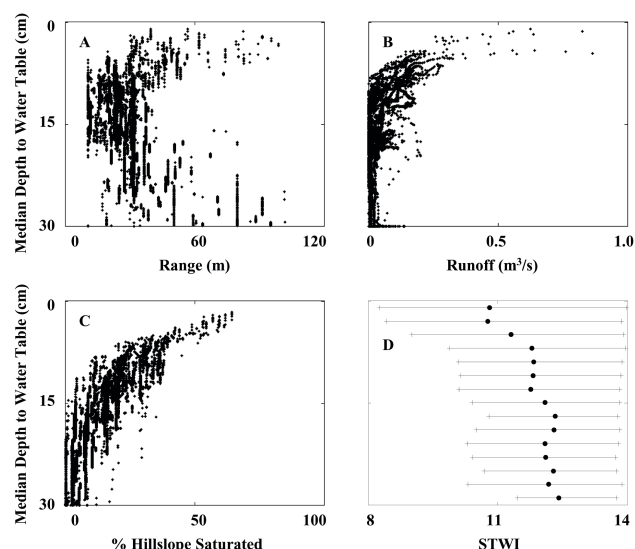


Fig. 4. From the short time interval data, variations in (A) the range [m], (B) runoff [m^3/s], (C) percentage of the hillslope saturated [%], and (D) average (black dots) STWI of the saturated area with bars showing minimum and maximum STWI of the saturated area with respect to the median depth to water table [cm] for the study site.

antecedent rainfall amounts greater than 1 cm in depth) and coincide with large rainfall amounts (rainfall events greater than 2 cm in depth). The median depth to water table fluctuated quickly and rose in response to rain events for the study site (Fig. 3C). The median water table was consistently close to the ground surface during March through early June. From mid-June through the end of August the median water table was deeper with high fluctuations during rain events (Fig. 3C). The water table and stream response to rainfall at the study site was typical for this region (Mehta et al., 2004; Agnew et al., 2006). High water tables near streams were maintained in spring (March through May) by interflow from either snowmelt or rainfall from upslope areas. The range parameter for the fitted exponential models was highly variable in time (Fig. 3D). The minimum range was about 9 m and the maximum 105 m.

From this time series, the importance the median water table plays in the hydrology of the hillslope was investigated. The ranges for the short time interval indicator semivariograms decreased as the median water table rose (Fig. 4A). This is due to transitioning from a smooth, continuous spatial structure to one that is more discontinuous. This trend changed when the median water table was about 10 cm deep. After this point, as the median water table rose closer to the soil surface the ranges began to increase. This increase in range was due to the spatial structure of the wettest sampling locations becoming more continuous over the field site. The runoff increased, as expected, when the median water table rose towards the soil surface (Fig. 4B). There was a large increase in runoff observed when the median water table was

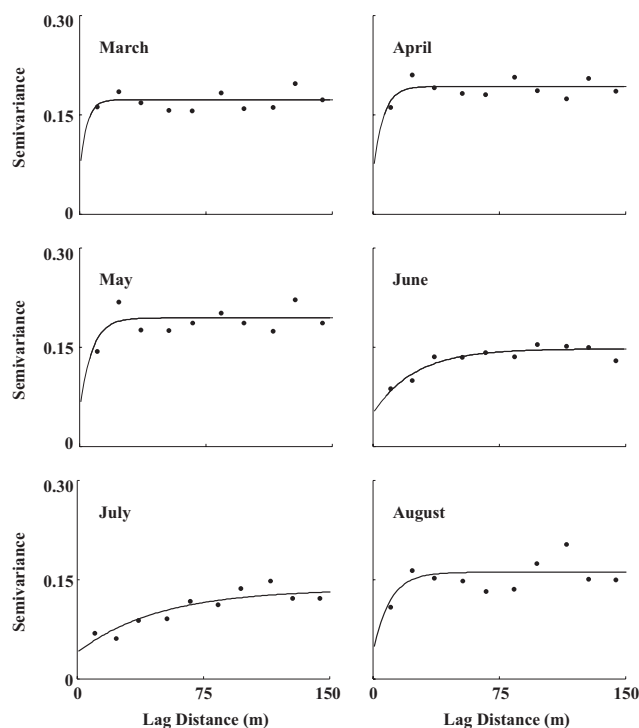


Fig. 5. Semivariograms for study site using the probability of exceeding the time-variable threshold (median water level) at each sampling location during the different months using long time interval analysis. The symbols are the sample semivariogram and the curves are the fitted exponential models.

closer to the soil surface than 10 cm. There was also an increase in the saturated portion of the hillslope with decrease in median depth to water table (Fig. 4C). For each 2-cm increment in water table depth the STWI values of all respective piezometers were grouped together and the mean, along with maximum and minimum, were computed. The mean STWI of all the saturated locations decreased as the median water table rose towards the soil surface (Fig. 4D). Also, the minimum STWI for all the sampling locations that saturate tended to decrease as the water table rose to the soil surface while the maximum STWI for all the sampling locations that saturate tended to stay constant.

4.2 Long time interval

The semivariograms for the long time interval data all had well defined sills and ranges (Fig. 5, Table 1). The nugget values for all months were similar ranging from 0.036 in August to 0.054 in March and April. Due to uncertainty associated with these nuggets, no further conclusions could be drawn from the nugget values. Sills from the semivariogram models corresponded to the average variance of the actual water table measures, which is expected. The sills varied from higher values during low median depth to water tables (0.193 in April, and 0.194 in May) to lower values

Table 1. Monthly characterization of the data set including median and variance of depth to water table, median and variance of frequency exceeding threshold, average and maximum runoff, daily average and total rainfall, and semivariogram parameters for the exponential models in Fig. 4 using the long time interval data for the study site.

Month	Depth to water table		Frequency exceeding threshold		Runoff		Rainfall		Semivariogram Parameters		
	Median [cm]	Variance [cm ²]	Median	Variance	Average [m ³ /s]	Max [m ³ /s]	Daily Average [cm]	Total [cm]	Nugget	Sill	Range [m]
March	16.60	91.7	0.089	0.140	0.047	0.20	0.69	15.2	0.054	0.173	12.0
April	13.05	110.2	0.162	0.174	0.031	0.29	0.65	19.6	0.054	0.193	17.3
May	15.75	118.5	0.123	0.165	0.025	0.87	0.54	16.7	0.048	0.194	20.9
June	31.85	71.9	0.082	0.066	0.033	0.22	0.30	9.1	0.049	0.147	73.8
July	35.25	53.5	0.028	0.024	0.011	0.83	0.59	18.3	0.039	0.135	144.4
August	16.51	90.6	0.106	0.092	0.002	0.01	0.10	2.3	0.036	0.161	29.0

Table 2. Summary of water table and rainfall for six dates used in IK analysis along with reduction in RMSE from cross validation with jackknifing between IK with hard data alone and IK with soft data.

Date	Median depth to water table [cm]	20th/80th percentile depth to water table [cm]	Variance of water table [cm ²]	1 day antecedent rainfall [cm]	RMSE		
					Reduction [%]	IK [%]	IK _{soft} [%]
27 March	8.4	2.9/16.6	61.6	1.1	4.3	46.3	44.3
2 April	6.2	1.3/12.8	59.5	1.5	11.7	28.2	24.9
13 April	7.7	2.6/19.2	72.6	2.7	8.5	61.5	56.3
26 April	5.8	1.0/12.7	46.7	2.4	1.2	24.8	24.5
3 May	6.4	0.5/17.2	93.8	2.1	9.9	46.7	42.1
26 May	4.4	0.4/10.5	41.6	3.6	7.9	24.0	22.1

during high median depth to water tables (0.147 in June and 0.135 in July). The ranges for the exponential models were longest in June and July at 73.8 and 144.4 m, respectively, and shorter during months with low median depth to water table. There was also noticeable difference in the relation between probability of exceeding the threshold and *STWI* for the long time interval (Fig. 6). For March through May, the linear regression between probability of exceeding the threshold and *STWI* had relatively higher slopes and positive intercepts compared to those for June through August.

The combined influence of long time interval and short time interval information on the spatial structure of the shallow water table was demonstrated visually for six rainfall events using kriging techniques (Fig. 7). These events were selected because they produced the highest median water tables for the period from March through May (i.e., when there was a noticeable increase in probability with increase in *STWI*) and characterized the hillslope response to rainfall for wet conditions (Table 2). For the 27 March and 3 May events, IK interpolations based on hard data alone showed high probability of exceeding the median water table in the

near stream region. Also, there was a region of high probability extending up the hillslope. Within this upslope region, there occurred discontinuous islands of higher probabilities (Fig. 7). Incorporating soft data based on seasonal trends of the water table into the IK interpolation reduced the occurrence of these isolated islands of high probabilities (Fig. 7). For the events on 2, 13, 26 April and 26 May, IK interpolations based on hard data alone gave relatively high probability of exceeding the threshold for the area closest to the stream, but little high probability in the region further upslope. Incorporating soft data, i.e., *STWI*, the near stream region having high probability of exceeding the median water table was larger. Also, the topographically converging region upslope from the stream was predicted as having higher probability of exceeding the threshold when incorporating the soft data than when using hard data alone.

The improvement in interpolation made by incorporating soft data was evaluated by jackknifing to cross validate the kriging interpolations. This method of cross validation tests a kriging interpolation by dividing the original dataset to produce a testing and a training dataset. Randomly, 30% (14 of

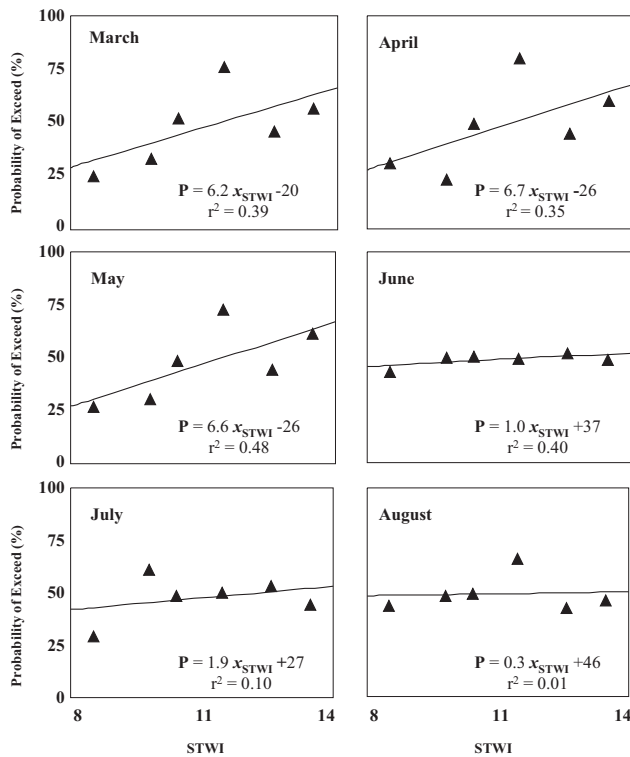


Fig. 6. Relationship between probability of exceeding the time-variable threshold (median water level) and STWI for each month. Points represent average probability of exceeding threshold and average STWI for each unit interval of STWI for the hillslope.

the 44 total) of the sampling locations were removed from the original dataset to create a testing dataset leaving 70% (30 of the 44 total) in a training dataset for analysis. To compare the interpolation methods, root mean square error (RMSE) was computed between the observed values in the testing dataset and predicted values using both methods. From these, the percentage reduction achieved by incorporating soft data evaluated. For each event, IK incorporating soft data reduces the RMSE for between the observed and predicted values (Table 2). This reduction in RMSE reflects a better representation of observed depths to water table using IK with soft data.

5 Discussion

Semivariograms based on the long time interval analysis (Fig. 5) demonstrate the seasonal controls on hydrology for this hillslope. There is a clear distinction between March through May and June through July. During wet conditions (i.e., March through May), shallow water tables in the convergent zones lead to shorter ranges in the long time interval semivariogram results. In drier conditions (i.e., June through July), the spatial structure of the more often wet sampling locations becomes smooth and homogenous in space. This is

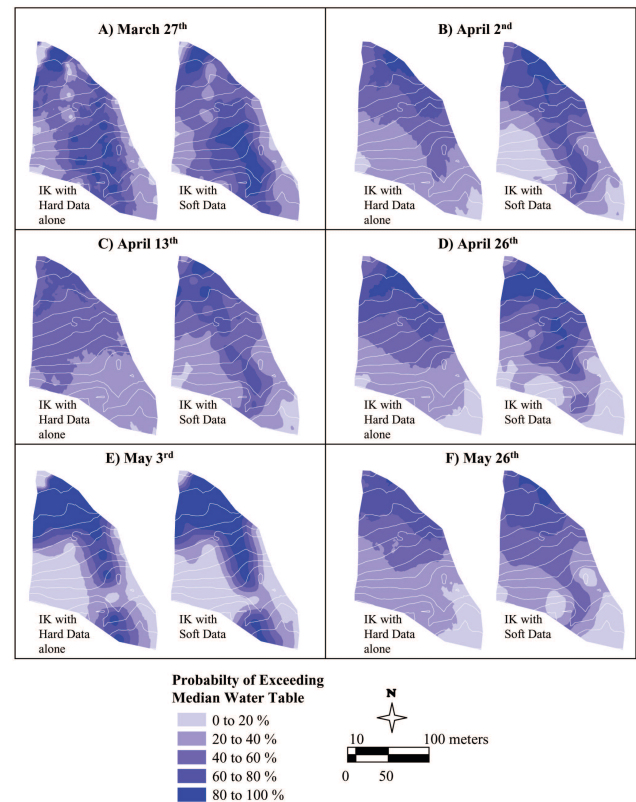


Fig. 7. IK with hard data alone and IK with soft data of study site for (A) 27 March, (B) 2 April, (C) 13 April, (D) 26 April, (E) 3 May, and (F) 26 May rain events using indicator values from short time interval for peak in rise of water table with 1 m contours as white lines.

seen in the semivariograms as longer ranges and lower sills. This is similar to the results of Western et al. (1998b) for soil moisture distributions in the Tarrawarra watershed. Locations where the water table is likely to rise during rainfall events have a shorter spatial structure during wet conditions than during dry where they constitute a more highly continuous spatial field behavior. Based on these semivariograms, August, which traditionally is a dry month, is between wet and dry conditions due to a large amount of rainfall occurring in late July and early August resulting from residual seasonal hurricane influence. This large amount of rainfall leads to a rise in the median water table not typical for the summer months. It is seen (Fig. 3C) that this high water table in August is not sustained and falls rapidly in periods of no rainfall. It lacks the snowmelt occurring from March through May that sustains the median water table close to the surface between rainfall events. The long-term analysis provides more information about prior conditions for the hillslope that describe the seasonal change in water table response as we move from the wet period to the dry period.

The short-term analysis, on the other hand, provides a way to describe changes in the spatial structure of the shallow

water table in response to rainfall events for this study site influenced by the antecedent conditions. During wet conditions, the local water table is close to the soil surface between rainfall events and, when rainfall occurs, the median water table raises producing surface saturation (Fig. 4C). Saturation causes increased ranges in indicator semivariograms (Fig. 4A). Since the expansion and contraction of saturated regions occurs quickly, the longer ranges seen during rain events in the short time interval analysis are not reflected in the monthly, frequency-based analysis (Figs. 4A and 5). Conversely, when there is a deep water table (i.e., high depth to water table), the short time interval analysis shows a decrease in range with rising water table. Rain falling during these deeper water table conditions permeates and runs laterally as interflow along the restrictive layer down-slope to accumulate in converging areas (Fig. 4D). Since the interflow is channeled into these converging areas that are relatively smaller than the total area and surface saturation tends not to occur, the spatial structure of the wettest locations becomes less homogeneous (discontinuous) compared with that of uniformly dry conditions.

Previous work demonstrated the link between antecedent conditions, in the form of pre-event depth to water table, and probability of saturation at the sampling locations during wet conditions (Lyon et al., 2006). The higher water tables are prerequisite for the lateral expansion of large-scale saturated source areas seen in the short time interval analysis. The lateral extent of expansion is captured with the decreasing minimum (and constant maximum) *STWI* for these saturated areas as the median water table approaches the soil surface (Fig. 4D). Thus, saturation for this hillslope during wet conditions starts at the highest *STWI* values and spreads to locations of lower *STWI*. *STWI* values are highest at locations that combine large upslope areas, low local slopes, and shallow soils. Since these saturated regions expand along gradients of decreasing *STWI*, the indicator semivariograms exhibit increased ranges for these rainfall events producing large, expanding saturated areas. This reaction is common to watersheds where saturation excess overland flow is considered a dominant pathway during wet season rain events (Western et al., 2004). This lateral expansion of saturated areas has been observed by other researchers in the Catskills due to accumulation of interflow water in the form of increased soil moisture at the hill bottoms relative to the steep parts of the hills during wet periods (Frankenberger et al., 1999; Ogden and Watts, 2000; Mehta et al., 2004). These studies observed locations where saturation commonly occurred are those where (1) the soil above the low conductivity layer is shallow, (2) the slope decreases downhill, such as the toe-slope of a hill, or (3) in topographically converging areas. In this study, occurrences of exceeding the median water table, which may be an adequate surrogate for saturation during high water table conditions, were observed at all three locations.

Without additional information such as provided by environmental traces it is not possible to discern the exact hydro-

drological pathways. Still, identifying spatial patterns of saturation is assumed to provide important information when focusing on topics such as non-point source pollution control (Walter et al., 2005). Throughout the observation period, increases in runoff, which were non-linear with respect to the median depth to water table, were observed when there were increased saturated areas (Fig. 4A). A possible interpretation is that as surface saturation regions expand, more rainfall is directly contributing to stream flow as overland flow or rapid subsurface flow. There seemed to be a threshold above which the median water table must rise before “runoff” to the stream increased dramatically (Fig. 4B). It is when the median water table raises above this threshold that longer ranges are observed in the indicator semivariograms due to expansion of surface saturated areas. The identification of these spatial patterns of saturation, which can be used to control where and when chemicals and nutrients are delivered, can be an important hydrological component for non-point source pollution control. Using kriging techniques (Fig. 7), the semivariogram analysis used to investigate the spatial and temporal evolution of the shallow water table can be further employed to identify such physical patterns on the hillslope.

The use of geostatistical techniques, such as IK, is influenced by the number of sampling locations. Western et al. (1998a) suggest that a large dataset is required to produce reliable results. For this study, since the water table was below the extent of the sampling devices over parts of the sampling period, traditional, measurement-based semivariogram analysis was not an option. By transforming measures into indicator values, water tables deeper than the sampling devices could still be included in the analysis. The limited number of sampling locations produced large fluctuations in the indicator semivariogram ranges for the short time interval analysis. This led to poor representations when kriging. However, the length of the sampling period has allowed for the use of soft data in combination with IK to create a more robust interpolation of the observed data that incorporated different timescales. This compensated for sparse spatial coverage and incorporated the seasonal variations in the hydrology of the region into the dataset. The *STWI* used in this study correlated well with probability of exceeding the median water table during the wet period. This is similar to results from wet periods for long-term modeling of this watershed (Agnew et al., 2006). When the median water table is close to the soil surface, such as in periods of snowmelt, the probability of exceeding the median water table coincides to the probability of saturation. The influence of topography during drier periods when the water table is not near the soil's surface, however, is not well established. For the wet period, the soft data created with prior probability allowed for IK that represented the physical process of the hillslope. This smoothed the kriging and eliminated islands of high probability of exceeding the threshold. These isolated regions are attributed to the sparse nature of the point observations and position of sampling sites influencing the IK. Using soft data,

the data are interpolated in a manner consistent with the underlying hydrologic processes for the hillslope to represent the influence of both event-based and seasonal trends.

By incorporating the soft data with the IK, the overall error in interpolation for the data was reduced. This provided better information about where on the hillslope hydrologically active areas occur. These regions may be extremely important in the development of nutrient management plans and in control the transport of pollutants. Also, by developing soft data based on readily available spatial data (i.e., DEM and SURRGO), prior probabilities could be developed from other analysis techniques if long time interval data such as those used in this study are not available. Lyon et al. (2006) improved interpolations by incorporating soft data into IK, but these improvements were limited to locations where the pre-event water table was known explicitly. Long-term modeling studies, such as that of Agnew et al. (2006), can provide the prior probability to create soft data for fortifying hard data observations. Thus, fewer observations can be made without compromising the robustness of the spatial data obtained. In addition, these soft data, when occurring at longer temporal scales, can provide information about seasonal variations in spatial patterns that heavily influence hydrology. Data based on interpolations of this style provide sources for validation of long-term risk assessment models. They can also aid in the development of appropriate techniques to better model saturated area formation by spatially representing data about water table response to rainfall events. Incorporation of soft data leads to a more realistic representation of hillslope reaction to rainfall events by including processes involved in the formation of saturated areas. This style of geostatistical analysis gives a manner to organize and represent spatial changes in the shallow ground water table. These changes occur at different temporal scales that can be integrated to better describe hillslope-scale hydrological processes.

6 Conclusions

Geostatistical methods were used to describe the spatial structure of the shallow water table in the near stream region. Using 44 sampling locations from a study site in the Town Brook watershed in the Catskill Mountain region of New York State, indicator methods have been used to explore variations in both short time intervals (15-min) and long time intervals (months). These time intervals were able to describe the event-based reaction of the shallow water table and the seasonal trends influencing the hydrology of the hillslope. The shallow water table for the study site shows two distinct responses depending on the position of the median water table. When the median water table was near to the soil surface (wet conditions), rainfall cause extensive surface saturation resulting in longer ranges in the indicator semivariograms. This is caused by expansion of saturated areas in topographically converging zones. During dry periods with deep water

tables, interflow concentrates the water table response causing decreases in range compared to the homogeneity in spatial structure prior to rainfall events. It was possible to visualize these changes in spatial structure using kriging techniques incorporating both the event-based and seasonal trends in the shallow water table response. This provided more realistic interpolations during high water table conditions by capturing structure in the shallow water table not available when using hard data alone. This type of kriging analysis provides a manner to locate physical patterns influencing the hydrology of the study site that are useful for validation of hydrological and contaminant transport modes. This study presents methods to characterize large amounts of point data temporally and spatially that can emphasize the hillslope-scale hydrological processes. By representing both spatial patterns and temporal evolution in the shallow water table with geostatistical analysis, saturated source areas active in controlling not only VSA runoff but also other hydrological pathways can be identified. Understanding this temporal evolution in the spatial structure of the shallow water table is the “where” and “when” of hydrology that is the groundwork for tasks such as non-point source pollution control.

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